

2002 INTERPRETIVE SUMMARY

NEW YORK CITIZENS STATEWIDE LAKE ASSESSMENT PROGRAM (CSLAP)

EAGLE LAKE

NY Federation of Lake Associations
NYS Department of Environmental Conservation

June, 2003

BACKGROUND AND ACKNOWLEDGMENT

The Citizens Statewide Lake Assessment Program (CSLAP) is a volunteer lake monitoring program conducted by the NYS Department of Environmental Conservation (NYSDEC) and the NYS Federation of Lake Associations (FOLA). Founded in 1986 with 25 pilot lakes, the program now involves more than 125 lakes, ponds, and reservoirs and 1000 volunteers from eastern Long Island to the Northern Adirondacks to the western-most lake in New York, including several Finger Lakes, Lake Ontario, and lakes within state parks. In this program, lay volunteers trained by the NYSDEC and FOLA collect water samples, observations, and perception data every other week in a fifteen-week interval between May and October. Water samples are analyzed by the NYS Department of Health and other certified laboratories. Analytical results are interpreted by the NYSDEC and FOLA, and utilized for a variety of purposes by the State of New York, local governments, researchers, and, most importantly, participating lake associations. This report summarizes the 2002 sampling results for **Eagle Lake**.

Eagle Lake is a 422 acre, class B lake found in the Town of Ticonderoga in Essex County, within the Eastern Adirondack region of New York State. Eagle Lake was first sampled as part of CSLAP in 2000. The following volunteers have participated in CSLAP, and deserve most of the credit for the success of this program at **Eagle Lake**: **Paul and Mary Lloyd Burroughs**.

In addition, the authors wish to acknowledge the following individuals, without whom this project and report would never have been completed:

From the Department of Environmental Conservation, N.G. Kaul, Sal Pagano, Dan Barolo, Italo Carcich, Phil DeGaetano, and Dick Draper, for supporting CSLAP for the past seventeen years; Jay Bloomfield and James Sutherland, for their work in developing and implementing the program; and the technical staff from the Lake Services Section, for continued technical review of program design.

From the Federation of Lake Associations, Anne Saltman, Dr. John Colgan, Don Keppel, Lew Stone, George Kelley, Nancy Mueller and the Board of Directors, for their continued strong support of CSLAP.

The New York State Department of Health (prior to 2002), particularly Jean White, and Upstate Freshwater Institute (in 2002), particularly Carol Matthews, provided laboratory materials and all analytical services, reviewed the raw data, and implemented the quality assurance/quality control program.

Finally, but most importantly, the authors would like to thank the more than 1000 volunteers who have made CSLAP a model for lay monitoring programs throughout the country and the recipient of a national environmental achievement award. Their time and effort have served to greatly expand the efforts of the state and the public to protect and enhance the magnificent water resources of New York State.

EAGLE LAKE FINDINGS AND EXECUTIVE SUMMARY

Eagle Lake was sampled as part of the New York Citizens Statewide Lake Assessment Program in 2002. For all program waters, water quality conditions and public perception of the lake each year and historically have been evaluated within annual reports issued after each sampling season. This report attempts to summarize both the 2002 CSLAP data and an historical comparison of the data collected within the 2002 sampling season and data collected at Eagle Lake prior to 2002.

The majority of the short- and long-term analyses of the water quality conditions in Eagle Lake are summarized in Table 2, divided into assessments of eutrophication indicators, other water quality indicators, and lake perception indicators. The 2002 data indicate that the lake continues to be best classified as oligotrophic, or highly unproductive. The lake experienced little change in the trophic indicators (clarity, phosphorus, chlorophyll *a*) measured through CSLAP, and the slight variability in these indicators (slightly lower clarity, nutrient and algae levels) is probably within the “normal” and expected range for Eagle Lake. The nitrogen to phosphorus ratios indicate that algae levels in Eagle Lake are controlled by phosphorus. There do not appear to be any significant seasonal water quality patterns, although deepwater nutrient levels are slightly elevated, indicating that this does not result in elevated surface nutrient levels after the lake turns over. Surface phosphorus levels in the lake have been consistently below the state phosphorus guidance value, and as a result, water transparency readings are consistently above the minimum recommended water clarity for swimming beaches. In short, water quality conditions indicated no significant changes in 2002.

The lake is weakly colored (low levels of dissolved organic matter) and it is likely that these readings reflect the soil and vegetation characteristics of the watershed (i.e. “natural” conditions at the lake). Color readings are probably not high enough to exert limits on the water transparency, even when algae levels are low. The lake has water of intermediate hardness, slightly alkaline (above neutral) pH readings, and undetectable nitrate readings. Conductivity readings have been stable since CSLAP sampling began in 2000. pH readings consistently fall within the NYS water quality standards (=6.5 to 8.5), and should not represent a problem for Eagle Lake. Nitrate and ammonia levels do not appear to warrant a threat to the lake, and the primary component of nitrogen appears to be organic (bound in algae cells).

The recreational suitability of Eagle Lake continues to be somewhat favorable. Recreational conditions in the lake have regularly been described as “slightly impaired” for most uses, due to “not quite crystal clear” conditions and weed growth to the lake surface, and excessive weed growth is often cited as impacting lake uses. These assessments were comparable to those measured in previous years, and are mostly stable over the course of a typical sampling season- this itself is consistent with the stability in water quality and weed densities over these periods. These assessments are typical of other lakes with moderate weed growth, but are much less favorable than in other lakes with comparable water quality characteristics.

The 1996 NYSDEC Priority Waterbody Listings (PWL) for the Upper Hudson River drainage basin do not include Eagle Lake. The CSLAP datasets suggest that *recreation* may be *stressed* due to excessive weed growth. The next PWL cycle for the Upper Hudson River drainage basin will occur in 2003.

I. INTRODUCTION: CSLAP DATA AND YOUR LAKE

Lakes are dynamic and complex ecosystems. They contain a variety of aquatic plants and animals that interact and live with each other in their aquatic setting. As water quality changes, so too will the plants and animals that live there and these changes in the food web also may additionally affect water quality. Water quality monitoring provides a window into the numerous and complex interactions of lakes. Even the most extensive and expensive monitoring program cannot **completely assess** a lake's water quality. However, by looking at some basic chemical, physical, and biological properties, it is possible to gain a greater understanding of the general condition of lakes. CSLAP monitoring is a basic step in overall water quality monitoring.

Understanding Trophic States

All lakes and ponds undergo **eutrophication**, an aging process, which involves stages of succession in biological productivity and water quality (see Figure 1). **Limnologists** (scientists who study fresh water systems) divide these stages into **trophic** states. Each trophic state can represent a wide range of biological, physical, and chemical characteristics and any lake may “naturally” be categorized within any of these trophic states. In general, the increase in productivity and decrease in clarity corresponds with an enrichment of nutrients, plant and animal life. Lakes with low biological productivity and high clarity are considered **oligotrophic**. Highly productive lakes with low clarity are considered **eutrophic**. Lakes that are **mesotrophic** have intermediate or moderate productivity and clarity. Eutrophication is a natural process, and is not necessarily indicative of man-made pollution.

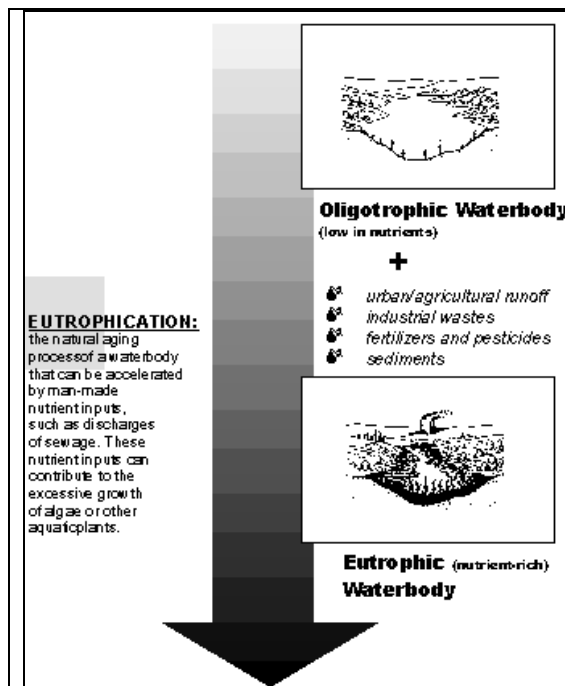


Figure 1. Trophic States

In fact, some lakes are thought to be “naturally” productive. It is important to understand that trophic classifications are not interchangeable with assessments of water quality. One person's opinion of degradation may be viewed by others as harmless or even beneficial. For example, a eutrophic lake may support an excellent warm-water fishery because it is nutrient rich, but a swimmer may describe that same lake as polluted. A lake's trophic state is still important because it provides lake managers with a reference point to view changes in a lake's water quality and begin to understand how these changes may cause **use impairments** (threaten the use of a lake or swimming, drinking water or fishing).

When human activities accelerate lake eutrophication, it is referred to as **cultural eutrophication**. Cultural eutrophication may result from shoreline erosion, agricultural and urban runoff, wastewater discharges or septic seepage, and other nonpoint source pollution sources. These can greatly accelerate the natural aging process of lakes, cause succession changes in the plant and animal life within the lake, shoreline and surrounding watershed, and impair the water quality and value of a lake. They may ultimately extend aquatic plants and emergent vegetation throughout the lake, resulting in the transformation of the lake into a marsh, prairie, and forest. The extent of cultural eutrophication,

and the corresponding pollution problems, can be signaled by significant changes in the trophic state over a short period of time.

II. CSLAP PARAMETERS

CSLAP monitors several parameters related to the trophic state of a lake, including how clear the water is, the amount of nutrients in the water, and the amount of algae growth resulting from those nutrients. Three parameters are the most important measures of eutrophication in most New York lakes: **total phosphorus**, **chlorophyll *a*** (measuring algal standing crop), and **Secchi disk transparency**. Because these parameters are closely linked to the growth of weeds and algae, they provide insight into “how the lake looks” and its suitability for recreation and aesthetics. Other CSLAP parameters help characterize water quality at the lake while balancing fiscal and logistic necessities. In addition, CSLAP also uses the responses on the **Field Observation Forms** to gauge volunteer perceptions of lake water quality. Most water quality “problems” arise from impairment of accepted or desired lake uses, or the perception that such uses are somehow degraded. As such, any water quality monitoring program should attempt to understand the link between perception and measurable quality.

The parameters analyzed in CSLAP provide valuable information for characterizing lakes. By adhering to a consistent sampling protocol provided in the CSLAP Sampling Protocol, volunteers collect and use data to assess both seasonal and yearly fluctuations in these parameters, and to evaluate the water quality in their lake. By comparing a specific year's data to historical water quality information, lake managers can pinpoint trends and determine if water quality is improving, degrading or remaining stable. Such a determination answers a first critical question posed in the lake management process.

Ranges for Parameters Assessing Trophic Status and Eagle Lake

The relationship between phosphorus, chlorophyll *a*, and Secchi disk transparency has been explored by many researchers, in hopes of assessing the trophic status (the degree of eutrophication) of lakes. Figure 3 shows ranges for phosphorus, chlorophyll *a*, and Secchi disk transparency (summer median) are representative for the major trophic classifications:

These classifications are valid for clear-water lakes only (waters with less than 30 platinum color units). Some humic or “tea color” lakes, for

Figure 2. Trophic Status Indicators

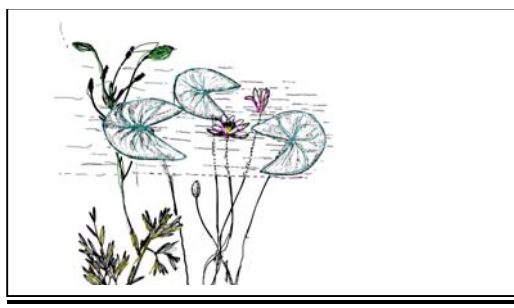
| Parameter | Eutrophic | Mesotrophic | Oligotrophic | Eagle Lake |
|-----------------------------|------------------|--------------------|---------------------|-------------------|
| Phosphorus (mg/l) | > 0.020 | 0.010 - 0.020 | < 0.010 | 0.007 |
| Chlorophyll <i>a</i> (µg/l) | > 8 | 2- 8 | < 2 | 1.5 |
| Secchi Disk Clarity (m) | 2 | 2- 5 | > 5 | 6.7 |

example, naturally have dissolved organic material with greater than 30 color units. This will cause the water transparency to be unexpectedly poor relative to low phosphorus and chlorophyll *a* levels. Water transparency can also be surprisingly lower than expected in shallow lakes, due to influences from the bottom. Even shallow lakes with high water clarity, low nutrient concentrations, and little algal growth may also have significant weed growth due to shallow water conditions. While such a lake may be considered unproductive by most standards, that same lake may experience severe aesthetic problems and recreational impairment related to weeds, not trophic state. Generally, however, the trophic relationships described above can be used as an accurate “first” gauge of productivity and overall water quality.

Figure 3. CSLAP Parameters

| PARAMETER | SIGNIFICANCE |
|---|---|
| Water Temperature (°C) | Water temperature affects many lake activities, including the rate of biological growth and the amount of dissolved oxygen. It also affects the length of the recreational season |
| Secchi Disk Transparency (m) | Determined by measuring the depth at which a black and white disk disappears from sight, the Secchi disk transparency estimates the clarity of the water. In lakes with low color and rooted macrophyte ("weed") levels, it is related to algal productivity |
| Conductivity (µmho/cm) | Specific conductance measures the electrical current that passes through water, and is used to estimate the number of ions (charged particles). It is somewhat related to both the hardness and alkalinity (acid-buffering capacity) of the water, and may influence the degree to which nutrients remain in the water. Generally, lakes with conductivity less than 100 µmho/cm are considered softwater, while conductivity readings above 300 µmho/cm are found in hardwater lakes. |
| pH | pH is a measure of the (free) hydrogen ion concentration in solution. Most clearwater lakes must maintain a pH between 6 and 9 to support most types of plant and animal life. Low pH waters (<7) are acidic, while high pH waters (>7) are basic |
| Color (true) (platinum color units) | The color of dissolved materials in water usually consists of organic matter, such as decaying macrophytes or other vegetation. It is not necessarily indicative of water quality, but may significantly influence water transparency or algae growth. Color in excess of 30 ptu indicate sufficient quantities of dissolved organic matter to affect clarity by imparting a tannic color to the water. |
| Phosphorus (total, mg/l) | Phosphorus is one of the major nutrients needed for plant growth. It is often considered the "limiting" nutrient in NYS lakes, for biological productivity is often limited if phosphorus inputs are limited. Nitrogen to phosphorus ratios of >10 generally indicate phosphorus limitation. Many lake management plans are centered around phosphorus controls. It is measured as total phosphorus (TP) |
| Nitrogen (nitrate, ammonia, and total (dissolved), mg/l) | Nitrogen is another nutrient necessary for plant growth, and can act as a limiting nutrient in some lakes, particularly in the spring and early summer. Nitrogen to phosphorus ratios < 7 generally indicate nitrogen limitation (for algae growth). For much of the sampling season, many CSLAP lakes have very low or undetectable levels of one or more forms of nitrogen. It is measured in CSLAP in three forms- nitrate/nitrite (NO _x), ammonia (NH _{3/4}), and total nitrogen (TN or TDN). |
| Chlorophyll a (µg/l) | The measurement of chlorophyll a, the primary photosynthetic pigment found in green plants, provides an estimate of phytoplankton (algal) productivity, which may be strongly influenced by phosphorus |
| Calcium (mg/l) | Calcium is a required nutrient for most aquatic fauna, and is required for the shell growth for zebra mussels and other aquatic organisms. It is naturally contributed to lakes from limestone deposits and is often strongly correlated with lake buffering capacity and conductivity. |

By each of the trophic criteria listed above, the lake would be classified as an **oligotrophic, or highly unproductive lake**.



III. AQUATIC PLANTS

Macrophytes:

Aquatic plants should be recognized for their contributions to lake beauty as well as providing food and shelter for other life in the lake. Emergent and floating plants such as water lilies floating on the lake surface may provide aesthetic appeal with their colorful flowers; sedges and cattails help to prevent shoreline erosion, and may provide food and cover for birds. Submergent plants like pondweeds and leafy waterweed harbor insects, provide nurseries for amphibians and fish, and provide food for birds and other animals. Those who enjoy fishing at the lake appreciate a diverse plant population.

Aquatic plants can be found throughout the *littoral zone*, the near-shore areas in which sufficient light reaches the lake bottom to promote photosynthesis. Plant growth in any particular part of the lake is a function of available light, nutrition and space, bottom substrate, wave action, and other factors. A large portion of aquatic vegetation consists of the microscopic algae referred to as phytoplankton; the other portion is the larger rooted plants called **macrophytes**.

Of particular concern to many lakefront residents and recreational users are the *non-indigenous macrophyte species* that can frequently dominate a native aquatic plant community and crowd out more beneficial species. The species may be introduced to a lake by waterfowl, but in most cases they are introduced by fragments or seedlings that remain on watercraft from already-infested lakes. Once introduced, these species have tenacious survival skills, crowding out, dominating and eventually aggressively overtaking the indigenous (native) plant communities. When this occurs, they interfere with recreational activities such as fishing, swimming or water-skiing. **These species need to be properly identified to be effectively managed.**

Non-native Invasive Macrophyte Species

Examples of **the common non-native invasive species found** in New York are:

- **Eurasian watermilfoil** (*Myriophyllum spicatum*)
- **Curly-leaf pondweed** (*Potamogeton crispus*)
- **Eurasian water chestnut** (*Trapa natans*)
- **Fanwort** (*Cabomba caroliniana*).

If these plants are not present, efforts should be made to continue protecting the lake from the introduction of these species.

Whether the role of the lake manager is to better understand the lake ecosystem or better manage the aquatic plant community, knowledge of plant distribution is paramount to the management process. There are many procedures available for assessing and monitoring aquatic vegetation. The CSLAP Sampling Protocol contains procedures for a “semi-quantitative” plant monitoring program. Volunteers collect plant specimens and provide field information and qualitative abundance estimates for an assessment of the macrophyte communities within critical areas of the lake. While these techniques are no substitute for professional plant surveys, they can help provide better information for lake managers. Lake associations planning to devote significant time and expenditures toward a plant management program are advised to pursue more extensive plant surveying activities.

Aquatic plant surveys have not been conducted through CSLAP at Eagle Lake, although the presence of Eurasian watermilfoil (*Myriophyllum spicatum*) has been verified by other sources.

The Other Kind of Aquatic Vegetation

Microscopic algae referred to as phytoplankton make up much of aquatic vegetation found in lakes. For this reason, and since phytoplankton are the primary producers of food (through photosynthesis) in lakes, they are the most important component of the complex food web that governs ecological interactions in lakes.

In a lake, phytoplankton communities are usually very diverse, and are comprised of hundreds of species having different requirements for nutrients, temperature and light. In many lakes, including those of New York, diatom populations are greatest in the spring, due to a competitive advantage in cooler water and relatively high levels of silica. In most lakes, however, diatom densities rarely reach nuisance portions in the spring. By the summer, green algae take advantage of warmer temperatures and greater amounts of nutrients (particularly nitrogen) in the warm water and often increase in density. These algae often grow in higher densities than do diatoms or most other species, although they are often not the types of algae most frequently implicated in noxious algae blooms. Later in the summer and in

the early fall, blue green algae, which possess the ability to utilize atmospheric nitrogen to provide this required nutrient, increase in response to higher phosphorus concentrations. This often happens right before turnover, or destratification in the fall. These algae are most often associated with taste and odor problems, bloom conditions, and the “spilled paint” slick that prompts the most complaints about algae. Each lake possesses a unique blend of algal communities, often varying in population size from year to year, and with differing species proportional in the entire population. The most common types range from the mentioned diatoms, green, and blue-green algae, to golden-brown algae to dinoflagellates and many others, dominating each lake community.

So how can this be evaluated through CSLAP? CSLAP does assess algal biomass through the chlorophyll *a* measurement. While algal differentiation is important, many CSLAP lake associations are primarily interested in “how much?”, not “what kind?”, and this is assessed through the chlorophyll *a* measurement. Phytoplankton communities have not been regularly identified and monitored through CSLAP, in part due to the cost and difficulty in analyzing samples, and in part due to the difficulty in using a one-time sample to assess long-term variability in lake conditions. A phytoplankton analysis may reflect a temporary, highly unstable and dynamic water quality condition.

In previous CSLAP sampling seasons, nearly all lakes were sampled once for phytoplankton identification, and since then some lakes have been sampled on one or more occasions. For these lakes, a summary of the most abundant phytoplankton species is included below. Algal species frequently associated with taste and odor problems are specifically noted in this table, although it should be mentioned that these samples, like all other water samples collected through CSLAP, come from near the center of the lake, a location not usually near water intakes or swimming beaches. Since algal communities can also be spatially quite variable, even a preponderance of taste and odor-causing species in the water samples might not necessarily translate to potable water intake or aesthetic impairments, although the threat of such an impairment might be duly noted in the “Considerations” section below.

Phytoplankton surveys have not been conducted through CSLAP at Eagle Lake.

IV. EAGLE LAKE CSLAP WATER QUALITY DATA

CSLAP is intended to provide the strong data base which will help lake associations understand lake conditions and foster sound lake protection and pollution prevention decisions. This individual lake summary for 2002 contains two forms of information. The **raw data** and **graphs** present a snapshot or glimpse of water quality conditions at each lake. They are based on (at most) eight sampling events during the summer. As lakes are sampled through CSLAP for a number of years, the database for each lake will expand, and assessments of lake conditions and water quality data become more accurate. For this reason, lakes new to CSLAP for only one year will not have information about annual trends.

Raw Data

Two “**data sets**” are provided below. The data presented in Table 1 include an annual summary of the minimum, maximum, and average for each of the CSLAP sampling parameters, including data from other sources for which sufficient quality assurance/quality control documentation is available for assessing the validity of the results. This data may be useful for comparing a certain data point perhaps for the current sampling year with historical data information. Table 2 includes more detailed summaries of the 2002 and historical data sets, including some evaluation of water quality trends, comparison against existing water quality standards, and whether 2002 represented a typical year.

Graphs

The second form of data analysis for your lake is presented in the form of **graphs**. These graphs are based on the raw data sets to represent a snapshot of water quality conditions at your lake. The more sampling that has been done on a particular lake, the more information that can be presented on the graph, and the more information you have to identify annual trends for your lake. For example, a lake that has been doing CSLAP monitoring consistently for five years will have a graph depicting five years worth of data, whereas a lake that has been doing CSLAP sampling for only one year may only have one. Therefore, it is important to consider the number of sampling years of information in addition to where the data points fall on a graph while trying to draw conclusions about annual trends. There are certain factors not accounted for in this report that lake managers should consider:

- **Local weather conditions** (high or low temperatures, rainfall, droughts or hurricanes). Due to delays in receiving meteorological data from NOAA stations within NYS, weather data are not included in these reports. It is certain that some of the variability reported below can be attributed more to weather patterns than to a “real” water trend or change. However, it is presumed that much of the sampling “noise” associated with weather is dampened over multiple years of data collection, and thus should not significantly influence the limited trend analyses provided for CSLAP lakes with longer and larger databases.
- **Sampling season and parameter limitations.** Because sampling is generally confined to June-September, this report does not look at CSLAP parameters during the winter and other seasons. Winter conditions can impact the usability and water quality of a lake conditions. In addition, there are other sampling parameters (fecal coliform, dissolved oxygen, etc.) that may be responsible for chemical and biological processes and changes in physical measurements (such as water clarity) and the perceived conditions in the lake. **The 2002 CSLAP report attempts to standardize some comparisons by limiting the evaluation to the summer recreational season and the most common sampling periods (mid-June through mid-September).**
- **Statistical analyses.** True assessments of water quality trends and comparison to other lakes involve rigid statistical analyses. Such analyses are generally beyond the scope of this program, in part due to limitations on the time available to summarize data from nearly 100 lakes in the five months from data receipt to next sampling season. This may be due in part to the inevitable inter-lake inconsistencies in sampling dates from year to year, and in part to the limited scope of monitoring. Where appropriate, some statistical summaries, utilizing both parametric and non-parametric statistics, have been provided within the report (primarily in Table 2).
- **Mean versus Median-** Much of the water quality summary data presented in this report is reported as the **mean**, or the average of all of the readings in the period in question (summer, annual, year to year). However, while mean remains one of the most useful, and often most powerful, ways to estimate the most typical reading for many of the measured water quality indicators, it is a less useful and perhaps misleading estimate when the data are not “normally” distributed (most common readings in the middle of the range of all readings, with readings less common toward the end of the range).

TABLE 1: CSLAP Data Summary for Eagle Lake

| Year | Min | Avg | Max | N | Parameter |
|----------------|--------------|--------------|---------------|-----------|---------------------|
| 2000-02 | 5.00 | 6.69 | 7.95 | 23 | CSLAP Zsd |
| 2002 | 5.50 | 6.39 | 7.05 | 8 | CSLAP Zsd |
| 2001 | 6.95 | 7.24 | 7.60 | 7 | CSLAP Zsd |
| 2000 | 5.00 | 6.51 | 7.95 | 8 | CSLAP Zsd |
| 1999 | 6.40 | 7.90 | 9.50 | 3 | LCI Zsd |
| | | | | | |
| Year | Min | Avg | Max | N | Parameter |
| 2000-02 | 0.002 | 0.007 | 0.012 | 23 | CSLAP Tot.P |
| 2002 | 0.002 | 0.006 | 0.009 | 8 | CSLAP Tot.P |
| 2002 | 0.002 | 0.007 | 0.013 | 8 | CSLAP Hypo TP |
| 2001 | 0.004 | 0.006 | 0.012 | 7 | CSLAP Tot.P |
| 2000 | 0.004 | 0.007 | 0.010 | 8 | CSLAP Tot.P |
| 1999 | 0.006 | 0.007 | 0.008 | 2 | LCI Tot.P |
| 1999 | 0.008 | 0.010 | 0.011 | 2 | LCI Hypo Tot.P |
| | | | | | |
| Year | Min | Avg | Max | N | Parameter |
| 2000-02 | 0.00 | 0.01 | 0.01 | 22 | CSLAP NO3 |
| 2002 | 0.00 | 0.00 | 0.01 | 8 | CSLAP NO3 |
| 2001 | 0.01 | 0.01 | 0.01 | 7 | CSLAP NO3 |
| 2000 | 0.01 | 0.01 | 0.01 | 7 | CSLAP NO3 |
| 1999 | 0.01 | 0.01 | 0.01 | 2 | LCI NO3 |
| | | | | | |
| Year | Min | Avg | Max | N | Parameter |
| 2002-02 | 0.01 | 0.03 | 0.05 | 8 | CSLAP NH4 |
| 2002 | 0.01 | 0.03 | 0.05 | 8 | CSLAP NH4 |
| | | | | | |
| Year | Min | Avg | Max | N | Parameter |
| 2002-02 | 0.33 | 0.48 | 0.74 | 8 | CSLAP TDN |
| 2002 | 0.33 | 0.48 | 0.74 | 8 | CSLAP TDN |
| | | | | | |
| Year | Min | Avg | Max | N | Parameter |
| 2002-02 | 46.00 | 86.29 | 172.97 | 8 | CSLAP TN/TP |
| 2002 | 46.00 | 86.29 | 172.97 | 8 | CSLAP TN/TP |
| | | | | | |
| Year | Min | Avg | Max | N | Parameter |
| 2000-02 | 2 | 6 | 16 | 22 | CSLAP TColor |
| 2002 | 2 | 7 | 16 | 8 | CSLAP TColor |
| 2001 | 3 | 5 | 8 | 7 | CSLAP TColor |
| 2000 | 3 | 5 | 8 | 7 | CSLAP TColor |
| | | | | | |
| Year | Min | Avg | Max | N | Parameter |
| 2000-02 | 6.71 | 7.51 | 8.15 | 22 | CSLAP pH |
| 2002 | 6.71 | 7.58 | 7.95 | 8 | CSLAP pH |
| 2001 | 6.80 | 7.48 | 7.94 | 7 | CSLAP pH |
| 2000 | 6.75 | 7.46 | 8.15 | 7 | CSLAP pH |
| 1999 | 7.30 | 7.50 | 7.70 | 2 | LCI pH |

DATA SOURCE KEY

| | |
|--|--|
| CSLAP | New York Citizens Statewide Lake Assessment Program |
| LCI | the NYSDEC Lake Classification and Inventory Survey conducted during the 1980s and again beginning in 1996 on select sets of lakes, typically 1 to 4x per year |
| DEC | other water quality data collected by the NYSDEC Divisions of Water and Fish and Wildlife, typically 1 to 2x in any give year |
| ALSC | the NYSDEC (and other partners) Adirondack Lake Survey Corporation study of more than 1500 Adirondack and Catskill lakes during the mid 1980s, typically 1 to 2x |
| ELS | USEPA's Eastern Lakes Survey, conducted in the fall of 1982, 1x |
| NES | USEPA's National Eutrophication Survey, conducted in 1972, 2 to 10x |
| EMAP | USEPA and US Dept. of Interior's Environmental Monitoring and Assessment Program conducted from 1990 to present, 1 to 2x in four year cycles |
| Additional data source codes are provided in the individual lake reports | |

CSLAP DATA KEY:

The following key defines column headings and parameter results for each sampling season:

| | |
|-------------------------|--|
| L Name | Lake name |
| Date | Date of sampling |
| Zbot | Depth of the lake at the sampling site, meters |
| Zsd | Secchi disk transparency, meters |
| Zsp | Depth of the sample, meters |
| TAir | Temp of Air, °C |
| TH2O | Temp of Water Sample, °C |
| TotP | Total Phosphorus as P, in mg/l (Hypo = bottom sample) |
| NO3 | Nitrate + Nitrite nitrogen as N, in mg/l |
| NH_{3/4} | Ammonia as N, in mg/l |
| TN-TDN | Total Nitrogen = NO _x + NH _{3/4} + organic nitrogen, as N, in mg/l |
| TP/TN | Phosphorus/Nitrogen ratios |
| Ca | Calcium, in mg/l |
| Tcolor | True color, as platinum color units |
| pH | (negative logarithm of hydrogen ion concentration), standard pH |
| Cond25 | Specific conductance corrected to 25°C, in µmho/cm |
| Chl.a | Chlorophyll a, in µg/l |
| QA | Survey question re: physical condition of lake: (1) crystal clear; (2) not quite crystal clear; (3) definite algae greenness; (4) high algae levels; and (5) severely high algae levels |
| QB | Survey question re: aquatic plant populations of lake: (1) none visible; (2) visible underwater; (3) visible at lake surface; (4) dense growth at lake surface; (5) dense growth completely covering the nearshore lake surface |
| QC | Survey question re: recreational suitability of lake: (1) couldn't be nicer; (2) very minor aesthetic problems but excellent for overall use; (3) slightly impaired; (4) substantially impaired, although lake can be used; (5) recreation impossible |
| QD | Survey question re: factors affecting answer QC: (1) poor water clarity; (2) excessive weeds; (3) too much algae/odor; (4) lake looks bad; (5) poor weather; (6) litter, surface debris, beached/floating material; (7) too many lake users (boats, jetskis, etc); (8) other |

TABLE 1: CSLAP Data Summary for Eagle Lake (cont)

| Year | Min | Avg | Max | N | Parameter |
|----------------|-------------|----------------|-------------|-----------|---------------------|
| 2000-02 | 117 | 139 | 153 | 22 | CSLAP Cond25 |
| 2002 | 140 | 142 | 147 | 8 | CSLAP Cond25 |
| 2001 | 117 | 139 | 153 | 7 | CSLAP Cond25 |
| 2000 | 129 | 136 | 139 | 7 | CSLAP Cond25 |
| 1999 | 140 | 140 | 140 | 2 | LCI Cond25 |
| | | | | | |
| Year | Min | Avg | Max | N | Parameter |
| 2002-02 | 0 | #DIV/0! | 0 | 0 | CSLAP Ca |
| 2002 | 0 | #DIV/0! | 0 | 0 | CSLAP Ca |
| | | | | | |
| Year | Min | Avg | Max | N | Parameter |
| 2000-02 | 0.42 | 1.48 | 8.20 | 17 | CSLAP Chl.a |
| 2002 | 0.75 | 1.03 | 1.23 | 6 | CSLAP Chl.a |
| 2001 | 0.77 | 1.01 | 1.51 | 4 | CSLAP Chl.a |
| 2000 | 0.42 | 2.13 | 8.20 | 7 | CSLAP Chl.a |
| 1999 | 1.07 | 1.43 | 2.29 | 4 | LCI Chl.a |
| | | | | | |
| Year | Min | Avg | Max | N | Parameter |
| 2000-02 | 2 | 2.0 | 2 | 19 | QA |
| 2002 | 2 | 2.0 | 2 | 8 | QA |
| 2001 | 2 | 2.0 | 2 | 3 | QA |
| 2000 | 2 | 2.0 | 2 | 8 | QA |
| | | | | | |
| Year | Min | Avg | Max | N | Parameter |
| 2000-02 | 1 | 2.7 | 3 | 19 | QB |
| 2002 | 2 | 2.8 | 3 | 8 | QB |
| 2001 | 3 | 3.0 | 3 | 3 | QB |
| 2000 | 1 | 2.6 | 3 | 8 | QB |
| | | | | | |
| Year | Min | Avg | Max | N | Parameter |
| 2000-02 | 2 | 2.9 | 3 | 19 | QC |
| 2002 | 2 | 2.9 | 3 | 8 | QC |
| 2001 | 3 | 3.0 | 3 | 3 | QC |
| 2000 | 3 | 3.0 | 3 | 8 | QC |

In particular, comparisons of one lake to another, such as comparisons within a particular basin, can be greatly affected by the spread of the data across the range of all readings. For example, the average phosphorus level of nine lakes with very low readings (say 10 µg/l) and one lake with very high readings (say 110 µg/l) could be much higher (in this case, 20 µg/l) than in the “typical lake” in this set of lakes (much closer to 10 µg/l). In this case, **median**, or the middle reading in the range, is probably the most accurate representation of “typical”.

This report will include the use of both mean and median to evaluate “central tendency”, or the most typical reading, for the indicator in question. In most cases, “mean” is used most often to estimate central tendency. However, where noted, “median” may also be used.

TABLE 2- Present Year and Historical Data Summaries for Eagle Lake*Eutrophication Indicators*

| Parameter | Year | Minimum | Average | Maximum |
|------------|-----------|---------|---------|---------|
| Zsd | 2002 | 5.50 | 6.39 | 7.05 |
| (meters) | All Years | 5.00 | 6.69 | 7.95 |
| | | | | |
| Parameter | Year | Minimum | Average | Maximum |
| Phosphorus | 2002 | 0.002 | 0.006 | 0.009 |
| (mg/l) | All Years | 0.002 | 0.007 | 0.012 |
| | | | | |
| Parameter | Year | Minimum | Average | Maximum |
| Chl.a | 2002 | 0.75 | 1.03 | 1.23 |
| (µg/l) | All Years | 0.42 | 1.49 | 8.20 |

| Parameter | Year | Was 2002 Clarity the Highest or Lowest on Record? | Was 2002 a Typical Year? | Trophic Category | Zsd Changing? | % Samples Violating DOH Beach Std?+ |
|------------|-----------|---|--------------------------|------------------|-----------------|--|
| Zsd | 2002 | Within Normal Range | Yes | Oligotrophic | No | 0 |
| (meters) | All Years | | | Oligotrophic | | 0 |
| | | | | | | |
| Parameter | Year | Was 2002 TP the Highest or Lowest on Record? | Was 2002 a Typical Year? | Trophic Category | TP Changing? | % Samples Exceeding TP Guidance Value+ |
| Phosphorus | 2002 | Lowest at Times | Yes | Oligotrophic | No | 0 |
| (mg/l) | All Years | | | Oligotrophic | | 0 |
| | | | | | | |
| Parameter | Year | Was 2002 Algae the Highest or Lowest on Record? | Was 2002 a Typical Year? | Trophic Category | Chl.a Changing? | |
| Chl.a | 2002 | Within Normal Range | Yes | Oligotrophic | No | |
| (µg/l) | All Years | | | Oligotrophic | | |

+ - Minimum allowable water clarity for siting a new NYS swimming beach = 1.2 meters

+ - NYS Total Phosphorus Guidance Value for Class B and Higher Lakes = 0.020 mg/l

-The 2002 CSLAP dataset indicates that water quality conditions in Eagle Lake were mostly comparable to those measured in previous sampling seasons. Water clarity readings were slightly lower than in the typical CSLAP sampling season and phosphorus and algae levels were also slightly lower, but all appeared to be within the normal range of variability for the lake. There continues to be only a weak correlation between algae and clarity and between algae and nutrients. However, it is likely that any lake management activities undertaken to maintain water transparency must necessarily address algae levels in and nutrient loading to the lake. None of the trophic indicators change significantly over the summer (while clarity increases and phosphorus and chlorophyll readings decrease, these seasonal changes are not statistically significant), and while deepwater nutrient levels are slightly higher than those at the lake surface, it does not appear that this results in higher surface nutrient levels during the recreational season for the lake. Phosphorus levels in Eagle Lake are consistently below the state guidance value for lakes used for contact recreation (swimming), and as a result, water clarity readings have consistently exceeded the minimum recommended water transparency for swimming beaches (= 1.2 meters). In short, water quality conditions were generally similar in 2002 to those measured in the typical sampling season at Eagle Lake.

TABLE 2- Present Year and Historical Data Summaries for Eagle Lake (cont)
Other Water Quality Indicators

| Parameter | Year | Minimum | Average | Maximum |
|------------------|-------------|----------------|----------------|----------------|
| Nitrate | 2002 | 0.00 | 0.00 | 0.01 |
| (mg/l) | All Years | 0.00 | 0.01 | 0.01 |
| | | | | |
| Parameter | Year | Minimum | Average | Maximum |
| Ammonia | 2002 | 0.01 | 0.03 | 0.05 |
| (mg/l) | All Years | 0.01 | 0.03 | 0.05 |
| | | | | |
| Parameter | Year | Minimum | Average | Maximum |
| TDN | 2002 | 0.33 | 0.48 | 0.74 |
| (mg/l) | All Years | 0.33 | 0.48 | 0.74 |
| | | | | |
| Parameter | Year | Minimum | Average | Maximum |
| True Color | 2002 | 2 | 7 | 16 |
| (ptu) | All Years | 2 | 6 | 16 |
| | | | | |
| Parameter | Year | Minimum | Average | Maximum |
| pH | 2002 | 6.71 | 7.58 | 7.95 |
| (std units) | All Years | 6.71 | 7.52 | 8.15 |
| | | | | |
| Parameter | Year | Minimum | Average | Maximum |
| Conductivity | 2002 | 140 | 142 | 147 |
| (µmho/cm) | All Years | 117 | 139 | 153 |
| | | | | |
| Parameter | Year | Minimum | Average | Maximum |
| Calcium | 2002 | | | |
| (mg/l) | All Years | | | |

***- These data indicate Eagle Lake is a weakly colored, alkaline (above neutral pH) lake with consistently undetectable nitrate levels and water of intermediate hardness. Color readings do not appear to limit water clarity, even when algal densities are low. Nitrogen levels, primarily organic nitrogen, are sufficiently high that it appears that phosphorus controls algae growth (nitrogen to phosphorus ratios regularly exceed 25), and overall nitrogen levels are low. Neither nitrate nor ammonia appear to represent a threat to water quality. Conductivity readings have varied only slightly and in a manner that does not appear to be statistically significant. pH readings fall within the NYS water quality standards (=6.5 to 8.5) during all sampling sessions, and these pH readings should continue to adequately support most aquatic organisms.**

TABLE 2- Present Year and Historical Data Summaries for Eagle Lake (cont)
Other Water Quality Indicators (cont)

| Parameter | Year | Was 2002 Nitrate the Highest or Lowest on Record? | Was 2002 a Typical Year? | Nitrate High? | Nitrate Changing? | % Samples Exceeding NO3 Standard | |
|--------------|-----------|--|--------------------------|-------------------|--------------------|---|--------------------------------|
| Nitrate | 2002 | Both Highest and Lowest at Times | Lower Than Normal | No | No | 0 | |
| (mg/l) | All Years | | | No | | 0 | |
| | | | | | | | |
| Parameter | Year | Was 2002 Ammonia the Highest or Lowest on Record? | Was 2002 a Typical Year? | Ammonia High? | Ammonia Changing? | % Samples Exceeding NH4 Standard+ | |
| Ammonia | 2002 | Both Highest and Lowest at Times | Yes | No | | 0 | |
| (mg/l) | All Years | | | No | | 0 | |
| | | | | | | | |
| Parameter | Year | Was 2002 TDN the Highest or Lowest on Record? | Was 2002 a Typical Year? | TDN High? | TDN Changing? | Ratios of TN/TP Indicate P or N Limitation? | |
| TDN | 2002 | Both Highest and Lowest at Times | Yes | No | | P Limitation | |
| (mg/l) | All Years | | | No | | P Limitation | |
| | | | | | | | |
| Parameter | Year | Was 2002 Color the Highest or Lowest on Record? | Was 2002 a Typical Year? | Colored Lake? | Color Changing? | | |
| True Color | 2002 | Both Highest and Lowest at Times | Yes | No | No | | |
| (ptu) | All Years | | | No | | | |
| | | | | | | | |
| Parameter | Year | Was 2002 pH the Highest or Lowest on Record? | Was 2002 a Typical Year? | Acceptable Range? | pH Changing? | % Samples > Upper pH Standard+ | % Samples < Lower pH Standard+ |
| pH | 2002 | Lowest at Times | Yes | Yes | No | 0 | 0 |
| (std units) | All Years | | | Yes | | 0 | 0 |
| | | | | | | | |
| Parameter | Year | Was 2002 Conductivity Highest or Lowest on Record? | Was 2002 a Typical Year? | Relative Hardness | Conduct. Changing? | | |
| Conductivity | 2002 | Within Normal Range | Yes | Intermediate | No | | |
| (µmho/cm) | All Years | | | | | | |
| | | | | | | | |
| Parameter | Year | Was 2002 Calcium Highest or Lowest on Record? | Was 2002 a Typical Year? | | Calcium Changing? | | |
| Calcium | 2002 | | | | | | |
| (mg/l) | All Years | | | | | | |

+ - NYS Nitrate standard = 10 mg/l

+ - NYS pH standard- 6.5 < acceptable pH < 8.5

TABLE 2- Present Year and Historical Data Summaries for Eagle Lake (cont)*Lake Perception Indicators (1= most favorable, 5= least favorable)*

| Parameter | Year | Minimum | Average | Maximum |
|--------------|-----------|---------|---------|---------|
| QA | 2002 | 2 | 2.0 | 2 |
| (Clarity) | All Years | 2 | 2.0 | 2 |
| | | | | |
| Parameter | Year | Minimum | Average | Maximum |
| QB | 2002 | 2 | 2.8 | 3 |
| (Plants) | All Years | 1 | 2.7 | 3 |
| | | | | |
| Parameter | Year | Minimum | Average | Maximum |
| QC | 2002 | 2 | 2.9 | 3 |
| (Recreation) | All Years | 2 | 2.9 | 3 |

| Parameter | Year | Was 2002 Clarity the Highest or Lowest on Record? | Was 2002 a Typical Year? | | Clarity Changed? |
|--------------|-----------|---|--------------------------|--|---------------------|
| QA | 2002 | Highest and Lowest | Yes | | No |
| (Clarity) | All Years | | | | |
| | | | | | |
| Parameter | Year | Was 2002 Weed Growth the Heaviest on Record? | Was 2002 a Typical Year? | | Weeds Changed? |
| QB | 2002 | Heaviest at Times | Yes | | No |
| (Plants) | All Years | | | | |
| | | | | | |
| Parameter | Year | Was 2002 Recreation the Best or Worst on Record? | Was 2002 a Typical Year? | | Recreation Changed? |
| QC | 2002 | Both Best and Worst at Times | Yes | | No |
| (Recreation) | All Years | | | | |

-Recreational assessments of Eagle Lake in 2002 were somewhat favorable, similar to those measured in previous sampling seasons. The recreational suitability of the lake has been consistently described as “slightly impaired” for most uses, coincident with lake conditions described as being “not quite crystal clear” and aquatic plants that typically reach the lake surface. These assessments are typical of other lakes with “moderate” weed densities, but much less favorable than in other lakes with similar water quality characteristics. The perceived physical condition of the lake (at all times “not quite crystal clear”) is also less favorable than in other lakes with similar water transparency readings, suggesting the latter (although very high) may be lower than historical readings. The recreational suitability of the lake appears to be somewhat insensitive to changes in water quality, but excessive aquatic plant (weed) growth is most often cited as impacting recreational uses of the lake. These assessments are stable over the course of the typical sampling season, consistent with the relative stability in water quality and weed growth (at least during the summer).

How Do the 2002 Seasonal Data Compare to Historical Seasonal Data?

Seasonal Comparison of Eutrophication and Lake Perception Indicators–2002 Sampling Season and in the Typical Sampling Season at Eagle Lake

Figures 4 and 5 compare data for the measured eutrophication parameters for Eagle Lake in 2002 and since CSLAP sampling began at Eagle Lake. Figures 6 and 7 compare volunteer perception responses over the same time periods.

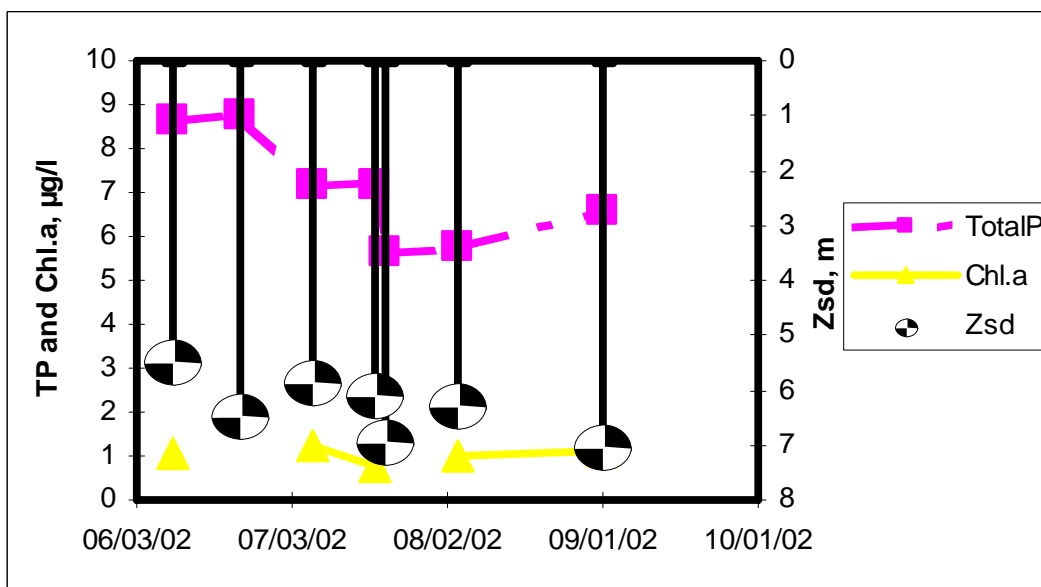


Figure 4. 2002 Eutrophication Data for Eagle Lake

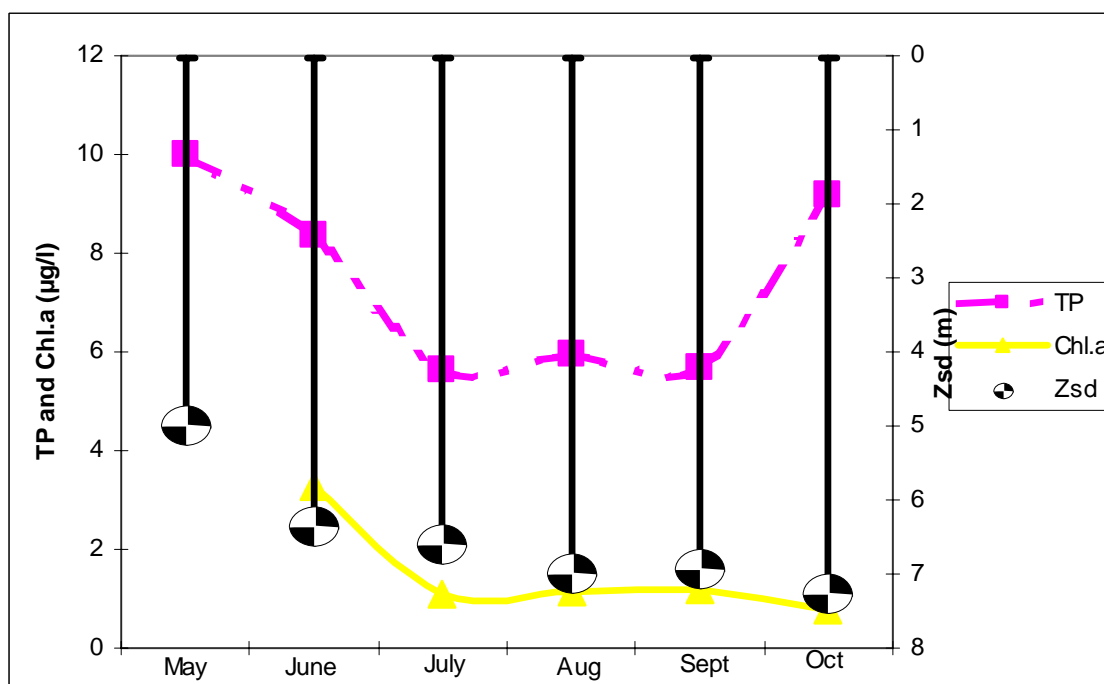


Figure 5- Eutrophication Data in a Typical (Monthly Mean) Year for Eagle Lake

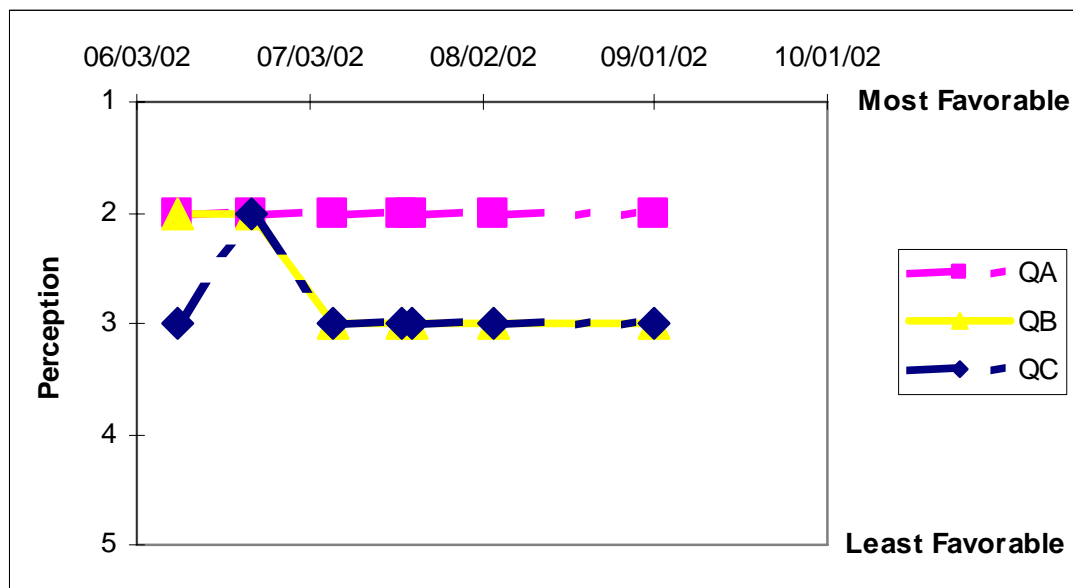


Figure 6. 2002 Lake Perception Data for Eagle Lake

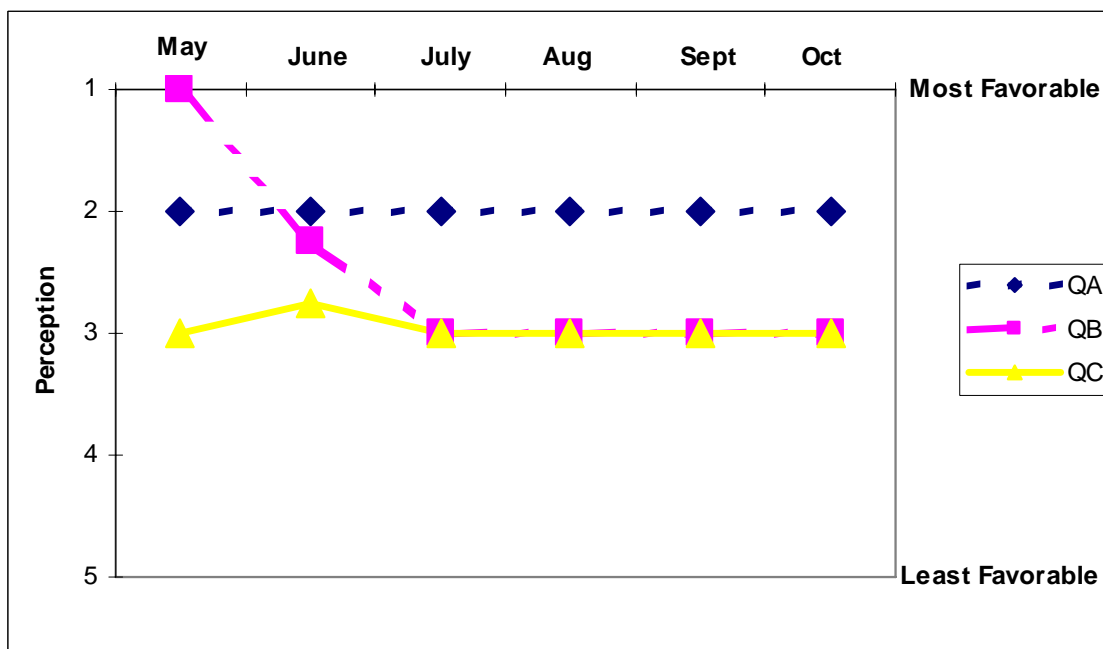


Figure 7- Lake Perception Data in a Typical (Monthly Mean) Year for Eagle Lake

(QA = clarity, ranging from (1) crystal clear to (3) definite algae greenness to (5) severely high algae levels
 QB = weeds, ranging from (1) not visible to (3) growing to the surface to (5) dense growth covers lake;
 QC = recreation, ranging from (1) could not be nicer to (3) slightly impaired to (5) lake not usable)

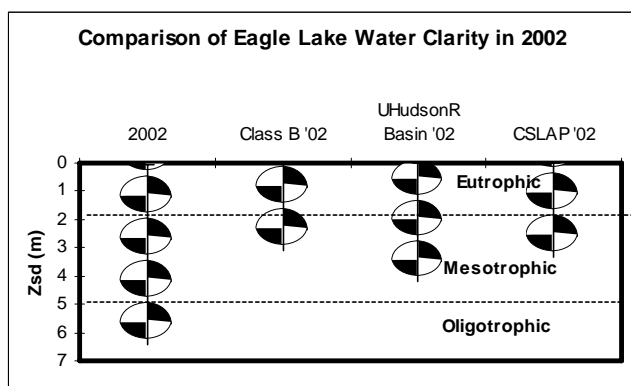


Figure 8. Comparison of 2002 Secchi Disk Transparency to Lakes With the Same Water Quality Classification, Neighboring Lakes, and Other CSLAP Lakes in 2002

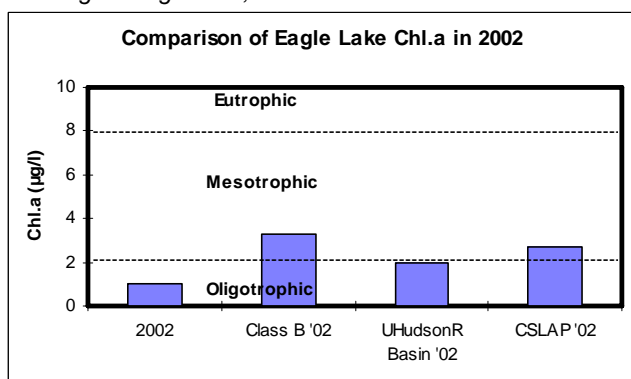


Figure 9. Comparison of 2002 Chlorophyll *a* to Lakes with the Same Water Quality Classification, Neighboring Lakes, and Other CSLAP Lakes in 2002

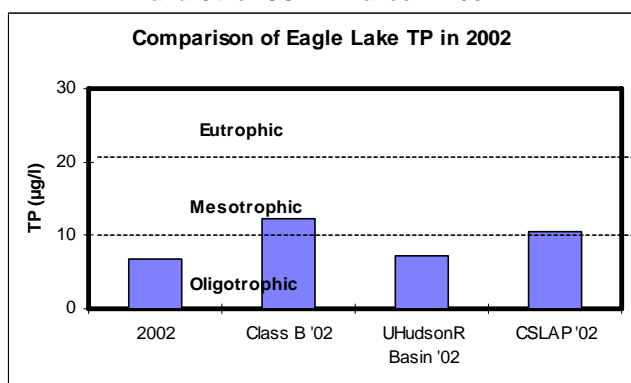


Figure 10. Comparison of 2002 Total Phosphorus to Lakes With the Same Water Quality Classification, Neighboring Lakes, and Other CSLAP Lakes in 2002

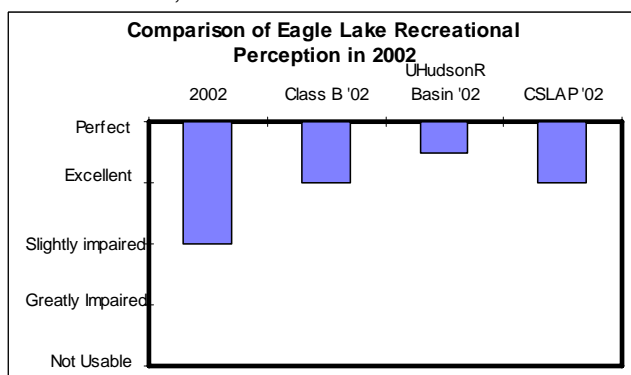


Figure 11. Comparison of 2002 Recreational Perception

How does Eagle Lake compare to other lakes?

Annual Comparison of Median Readings for Eutrophication Parameters and Recreational Assessment For Eagle Lake in 2002, Neighboring Lakes, Lakes with the Same Lake Classification, and Other CSLAP Lakes

The graphs to the left illustrate comparisons of each eutrophication parameter and recreational perception at Eagle Lake-in 2002, other lakes in the same drainage basin, lakes with the same water quality classification (each classification is summarized in Appendix B), and all of CSLAP. Please keep in mind that differences in watershed types, activities, lake history and other factors may result in differing water quality conditions at your lake relative to other nearby lakes. In addition, the limited data base for some regions of the state preclude a comprehensive comparison to neighboring lakes.

Based on these graphs, the following conclusions can be made about Eagle Lake in 2002:

- Using water clarity as an indicator, Eagle Lake was less productive than other lakes with the same water quality classification (Class B), other Upper Hudson River drainage basin lakes, and other CSLAP lakes.
- Using chlorophyll *a* concentrations as an indicator, Eagle Lake was less productive than other Class B, other Upper Hudson River drainage basin, and other CSLAP lakes.
- Using total phosphorus concentrations as an indicator, Eagle Lake was less productive than other Class B and other CSLAP lakes, and about as productive as other Upper Hudson River drainage basin lakes.
- Using QC on the field observations form as an indicator, Eagle Lake was less suitable for recreation than other Class B lakes, other Upper Hudson River drainage basin lakes and other CSLAP lakes.

V: PRIORITY WATERBODY AND IMPAIRED WATERS LIST

The Priority Waterbody List (PWL) is presently an inventory of all waters in New York State known to have designated water uses with some degree of impairment of which are threatened by potential impairment. However, the PWL is slowly evolving into an inventory of all waterbodies for which sufficient information is available to assess the condition and/or usability of the waterbody. PWL waters are identified through a broad network of county and state agencies, with significant public outreach and input, and the list is maintained and compiled by the NYSDEC Division of Water. Monitoring data from a variety of sources, including CSLAP, have been utilized by state and agencies to evaluate lakes for inclusion on the PWL, and the process for incorporating lakes data has become more standardized.

Specific numeric criteria have recently been developed to characterize sampled lakes in the available use-based PWL categories (precluded, impaired, stressed, or threatened). Evaluations utilize the NYS phosphorus guidance value, water quality standards, criteria utilized by other states, and the trophic ranges described earlier to supplement the other more antidotal inputs to the listing. The procedures by which waterbodies are evaluated are known as the Consolidated Assessment and Listing Methodology (CALM) process. This process is undertaken on an annual rotating basin, with waterbodies in several drainage basins evaluated each year. Each of the 17 drainage basins in the state are assessed within every five years.

Lakes that have been identified as precluded or impaired on the PWL are likely candidates for the federal 303(d) list, an “Impaired Waters” designation mandated by the federal Clean Water Act. Lakes on this list must be closely evaluated for the causes and sources of these problems. Remedial measures must be undertaken, under a defined schedule, to solve these water quality problems. This entire evaluation and remediation process is known as the “TMDL” process, which refers to the Total Maximum Daily Load calculations necessary to determine how much (pollution that causes the water quality problems) is too much.

TABLE 3- Water Quality Standards Associated With Class B and Higher Lakes

| <u>Parameter</u> | <u>Acceptable Level</u> | <u>To Protect.....</u> |
|--------------------------|-----------------------------|------------------------|
| Secchi Disk Transparency | > 1.2 meters* | Swimming |
| Total Phosphorus | < 0.020 mg/L and Narrative* | Swimming |
| Chlorophyll a | none | NA |
| Nitrate Nitrogen | < 10 mg/L and Narrative* | Drinking Water |
| Ammonia Nitrogen | 2 mg/L* | Drinking Water |
| True Color | Narrative* | Swimming |
| pH | < 8.5 and > 6.5* | Aquatic Life |
| Conductivity | None | NA |

*- Narrative Standards and Notes:

Secchi Disk Transparency: The 1.2 meter (4 feet) guidance is applied for safety reasons (to see submerged swimmers or bottom debris), and strictly applies only to citing new swimming beaches, but may be appropriate for all waterbodies used for contact recreation (swimming)

Phosphorus and Nitrogen: “None in amounts that will result in the growths of algae, weeds and slimes that will impair the waters for their best usages” (Class B= swimming)

-The 0.020 mg/l threshold for TP corresponds to a guidance value, not standard; it strictly applies to Class B and higher waters, but may be appropriate for other waterbodies used for contact recreation (swimming). NYS (and the other states) are in the process of identifying numerical nutrient (phosphorus, and perhaps Secchi disk transparency, chlorophyll *a*, and nitrogen) standards, but this is unlikely to be finalized within the next several years.

-The 10 mg/L Nitrate standard strictly applies to only Class A or higher waters, but is included here since some Class B lakes are informally used for potable water intake.

-For the form of ammonia (NH₃+NH₄) analyzed, a 2 mg/l human health standard applies to Class A or higher waters; while lower un-ionized ammonia standards apply to all classes of NYS lakes, this form is not analyzed through CSLAP

Color: “None in amounts that will adversely affect the color or impair the waters for their best usages” (for Class B waters, this is swimming)

pH: The standard applies to all classes of waterbodies

pH readings were within the NYS water quality standards (=6.5 to 8.5) during each of the CSLAP sampling sessions at Eagle Lake. Phosphorus levels at Eagle Lake have been well below the phosphorus guidance value for NYS lakes (=0.020 mg/l) during each of the CSLAP sampling sessions, and as a result, water transparency readings have at all times been well above the minimum recommended water clarity for swimming beaches (= 1.2 meters). It is not known if any of the narrative water quality standards listed in Table 3 have been violated at Eagle Lake.

Eagle Lake is not presently among the lakes listed on the Upper Hudson River drainage basin PWL (1996). The CSLAP dataset, including water chemistry data, physical measurements, and volunteer samplers’ perception data, indicate that *recreation* may be *stressed* by excessive weeds. The next PWL listing cycle for the Upper Hudson River drainage basin will occur in 2003.

GENERAL CONSIDERATIONS FOR ALL CSLAP LAKES

Nutrient controls can take several forms, depending on the original source of the nutrients:

- Septic systems can be regularly pumped or upgraded to reduce the stress on the leach fields which can be replaced with new soil or moving the discharge from the septic tank to a new field). Pumpout programs are usually quite inexpensive, particularly when lakefront residents negotiate a bulk rate discount with local pumping companies. Upgrading systems can be expensive, but may be necessary to handle the increased loading from camp expansion or conversion to year-round residency. Replacing leach fields alone can be expensive and limited by local soil or slope conditions, but may be the only way to reduce actual nutrient loading from septic systems to the lake. It should be noted that upgrading or replacing the leach field may do little to change any bacterial loading to the lake, since bacteria are controlled primarily within the septic tank, not the leach field.
- Stormwater runoff control plans include street cleaning, artificial marshes, sedimentation basins, runoff conveyance systems, and other strategies aimed at minimizing or intercepting pollutant discharge from impervious surfaces. The NYSDEC has developed a guide called Reducing the Impacts of Stormwater Runoff to provide more detailed information about developing a stormwater management plan. This is a strategy that cannot generally be tackled by an individual homeowner, but rather requires the effort and cooperation of lake residents and municipal officials.
- There are numerous agriculture management practices such as fertilizer controls, soil erosion practices, and control of animal wastes, which either reduce nutrient export or retain particles lost from agricultural fields. These practices are frequently employed in cooperation with county Soil and Water Conservation District offices, and are described in greater detail in the NYSDEC's Controlling Agricultural Nonpoint Source Water Pollution in New York State. Like stormwater controls, these require the cooperation of many watershed partners, including farmers.
- Streambank erosion can be caused by increased flow due to poorly managed urban areas, agricultural fields, construction sites, and deforested areas, or it may simply come from repetitive flow over disturbed streambanks. Control strategies may involve streambank stabilization, detention basins, revegetation, and water diversion.

Land use restrictions development and zoning tools such as floodplain management, master planning to allow for development clusters in more tolerant areas in the watershed and protection of more sensitive areas; deed or contracts which limit access to the lake, and cutting restrictions can be used to reduce pollutant loading to lakes. This approach varies greatly from one community to the next and frequently involves balancing lake use protection with land use restrictions. State law gives great latitude to local government in developing land use plans.

Lawn fertilizers frequently contain phosphorus, even though nitrogen is more likely to be the limiting nutrient for grasses and other terrestrial plants. By using lawn fertilizers with little or no phosphorus, eliminating lawn fertilizers or using lake water as a “fertilizer” at shoreline properties, fewer nutrients may enter the lake. Retaining the original flora as much as possible, or planting a buffer strip (trees, bushes, shrubs) along the shoreline, can reduce the nutrient load leaving a residential lawn.

Waterfowl introduce nutrients, plant fragments, and bacteria to the lake water through their feces. Feeding the waterfowl encourages congregation which in turn concentrates and

increases this nutrient source, and will increase the likelihood that plant fragments, particularly from Eurasian watermilfoil and other plants that easily fragment and reproduce through small fragments, can be introduced to a previously uncolonized lake.

Although not really a “watershed control strategy”, establishing **no-wake zones** can reduce shoreline erosion and local turbidity. Wave action, which can disturb flocculent bottom sediments and unconsolidated shoreline terrain is ultimately reduced, minimizing the spread of fertile soils to susceptible portions of the lake.

Do not discard or introduce plants from one water source to another, or deliberately introduce a "new" species from catalogue or vendor. For example, do not empty bilge or bait bucket water from another lake upon arrival at another lake, for this may contain traces of exotic plants or animals. Do not empty aquaria wastewater or plants to the lake.

Boat propellers are a major mode of transport to uncolonized lakes. Propellers, hitches, and trailers frequently get entangled by weeds and weed fragments. Boats not cleaned of fragments after leaving a colonized lake may introduce plant fragments to another location. New introductions of plants are often found near public access sites.

SPECIFIC CONSIDERATIONS FOR EAGLE LAKE

Management Focus: Water Clarity/Algae/Physical Condition/Recreational Condition

| Issue | Through | By? |
|------------------------|--------------------------------------|---|
| Maintain water clarity | Maintaining or reducing algae levels | Maintaining or reducing nutrient Inputs to the lake |

Discussion:

User perception and water quality data indicate a favorable physical condition and water clarity of the lake. This places the focus of water clarity management on maintaining present conditions, an enviable position for many other lake associations. Although some increase in nutrient loading is inevitable, the lake association should devote efforts to minimize the input of nutrients to the lake, or change activities that otherwise influence water clarity.

Management Focus: The Impact of Weeds on Recreational Condition

| Problem | Probable Cause | Probable Source |
|-----------------------------------|---|---|
| Moderate to Excessive weed growth | Shallow water depth, excessive nutrients and sediment | Excessive pollutant loading from watershed runoff (stormwater, construction sites, agriculture, etc.), septic, bottom disturbance,... |

Discussion:

Perception data indicate that aquatic weed growth is perceived to inhibit recreational use of this lake, at least in some parts of the lake or during certain times of the year. Nuisance weed growth in lakes is influenced by a variety of factors- water clarity, sediment characteristics, wave action, competition between individual plant species, sediment nutrient levels, etc. In most cases, excessive weed growth is associated with the presence of exotic, (non-native) submergent plant species such as Eurasian watermilfoil (*Myriophyllum spicatum*), although some lakes are inhibited by dense growth of native species. Some of these factors cannot be controlled by lake association activities, while others can only be addressed peripherally. For

example, sediment characteristics can be influenced by the solids loading to the lake. With the exception of some hand harvesting activities, aquatic plant management should only be undertaken when lake uses (recreational, municipal, economic, etc.) are significantly and regularly threatened or impaired. Management strategies can be costly and controversial, and a variety of factors should be weighed. Aquatic plant management most efficiently involves a mix of immediate, in-lake controls, and long-term measures to address the causes and sources of this excessive weed growth.

THE EAGLE LAKE ASSOCIATION HAS BEEN HEAVILY INVOLVED IN LOCAL AND STATEWIDE AQUATIC PLANT MANAGEMENT PLANNING, AND HAS BEEN DIRECTING LOCAL EFFORTS TO CONTROL INVASIVE EURASIAN WATERMILFOIL GROWTH WITH THE USE OF AQUATIC HERBICIDES (FLURIDONE). HOWEVER, THE ENTIRE RANGE OF AQUATIC PLANT MANAGEMENT TECHNIQUES ARE PRESENTED HERE FOR THE SAKE OF COMPLETENESS.

IN-LAKE CONTROL TECHNIQUES

The following strategies primarily address the cause, but not the ultimate source, of problems related to nuisance aquatic plant growth. As such, their effectiveness is necessarily short-term, but perhaps more immediately realized, than strategies that control the source of the problem. Until the sources of the problem are addressed, however, it is likely that these strategies will need to be continuously employed. Some of these are listed in the **Watershed Controls**, since many of the same pollutants contribute to excessive algae growth as well as nuisance weed growth. Except where noted, most of these in-lake techniques do not require permits in most parts of the state, but, as always, the NYDEC Region 5 Offices and the Adirondack Park Agency should be consulted before undertaking these strategies. These techniques are presented within the context of potential management for the conditions (types of nuisance plants, extent of problem) reported through CSLAP. In-lake control methods include: *physical/mechanical plant management techniques, chemical plant management techniques, and biological plant management techniques*

Physical/mechanical control techniques utilize several modes of operation to remove or reduce the growth of nuisance plants. The most commonly employed procedures are the following:

- *Mechanical harvesters* physically remove rooted aquatic plants by using a mechanical machine to cut and transport plants to the shore for proper storage. Mechanical harvesters are probably the most common “formal” plant management strategy in New York State. While it is essentially akin to “mowing the (lake) lawn”, it usually provides access to the lake surface and may remove some lake nutrients if the cut plants are disposed out of the watershed. However, if some shallow areas of the lake are not infested with weeds, they will likely become infested after mechanical harvesting, since fragments frequently wander from cut areas to barren sediment and colonize new plant communities. Harvesters are very expensive, but can be rented or leased. *Rotovators* are rotovating mechanical harvesters, dislodging and removing plants and roots. *Mechanical cutters* cut, but don’t remove, vegetation or fragments. Box springs, sickles, cutting bars, boat props, and anchors often serve as mechanical cutters.
- *Hand harvesting* is the fancy term for lake weeding- pulling out weeds and the root structure by hand. It is very labor intensive, but very plant selective (pull the “weeds”,

leave the “plants”); and can be effective if the entire plant is pulled and if the growth area is small enough to be fully cleared of the plant. *Diver dredging* is like hand harvesting with a vacuum cleaner- in this strategy, scuba divers hand-pull plants and place them into a suction hose for removal into a basket in a floating barge. It is also labor intensive and can be quite expensive, but it can be used in water deeper than about 5ft (the rough limit for hand harvesting). It works best where plant beds are dense, but is not very efficient when plant beds or stems are scattered.

- *Water level manipulation* is the same thing as *drawdown*, in which the lake surface is lowered, usually over the winter, to expose vegetation and sediments to freezing and drying conditions. Over time this affects the growing characteristics of the plants, and in many cases selectively eliminates susceptible plants. This is obviously limited to lakes that have a mechanism (dam structure, controlled culvert, etc.) for manipulating water level. It is usually very inexpensive, but doesn’t work on all plants and there is a risk of insufficient lake refill the following spring (causing docks to be orphaned from the waterfront). **It is not believed by the report authors that Eagle Lake can be sufficiently drawn down to utilize this technique.**
- *Bottom barriers* are screens or mats that are placed directly on the lake bottom to prevent the growth of weeds by eliminating sunlight needed for plant survival. The mats are held in place by anchors or stakes, and must be periodically cleaned or removed to detach any surface sediment that may serve as a medium for new growth. The mats, if installed properly, are almost always effective, with relatively few environmental side-effects, but are expensive and do not select for plant control under the mats. It is best used when plant communities are dense but small in area, and is not very efficient for lake-wide control.
- *Sediment removal*, also referred to as dredging, controls aquatic plants by physically removing vegetation and by increasing the depth of the lake so that plant growth is limited by light availability. Dredging projects are usually very successful at increasing depth and controlling vegetation, but they are very expensive, may result in significant side effects (turbidity, algal blooms, potential suspension of toxic materials), and may require significant area for disposal. This procedure usually triggers an extensive permitting process, **particularly in the Adirondack Park.**

Chemical control techniques involve the use of aquatic herbicides to kill undesired aquatic vegetation and prevent future nuisance weed growth. These herbicides come in granular or liquid formulations, and can be applied in spot- or whole-lake treatments. Some herbicides provide plant control by disrupting part of the plants life cycle or ability to produce food, while others have more toxicological effects. Aquatic herbicides are usually effective at controlling plants, but other factors in considering this option include the long term control (longevity), efficiency, and plant selectivity. Effectiveness may also depend on dosage rate, extent of non-target (usually native) plant growth, flushing rate, and other factors. The use of herbicides is often a highly controversial matter frequently influenced by personal philosophies about introducing chemicals to lakes. Some of the more recently registered herbicides appear to be more selective and have fewer side effects than some of the previously utilized chemicals. Chemical control of nuisance plants can be quite expensive, and, with only few exceptions, require permits and licensed applicators. **As discussed above, herbicides appear to be the control strategy of choice, at least among the active lake association members, at Eagle Lake.**

Biological control techniques presently involve the stocking of sterile grass carp, which are herbivorous fish that feed exclusively on macrophytes (and macroalgae). Grass carp, when stocked at the appropriate rate, have been effective at controlling nuisance weeds in many southern states, although their track record in NYS is relatively short, particularly in lakes with shallow or adjacent wetlands or in larger (>100 acre) lakes. These carp may not prefer the nuisance plant species desired for control (in particular Eurasian watermilfoil), and they are quite efficient at converting macrophyte biomass into nutrients that become available for algae growth. This is, however, one of the less expensive means of plant control. **The permitting process for grass carp in the Adirondacks is extensive.**

Naturally occurring biological controls may include native species of *aquatic weevils and moths* which burrow into and ultimately destroy some weeds. These organisms feed on Eurasian watermilfoil, and control nuisance plants in some Finger Lakes and throughout the Northeast. However, they also inhabit other lakes with varied or undocumented effectiveness for the long term. Because these organisms live in the canopy of weed beds and feed primarily on the top of the plants, harvesting may have a severe negative impact on the population. Research continues about their natural occurrence, and their effectiveness both as a natural or deliberately- introduced control mechanism for Eurasian watermilfoil. **The impact of herbivorous insects on Eurasian watermilfoil in Eagle Lake continues to be evaluated.**

Appendix A. Raw Data for Eagle Lake

| LNum | PName | Date | Zbot | Zsd | Zsamp | Tot.P | NO3 | NH4 | TDN | TN/TP | TColor | pH | Cond25 | Ca | Chl.a |
|------|---------|------------|------|------|-------|-------|------|------|------|--------|--------|------|--------|----|-------|
| 169 | Eagle L | 5/30/2000 | 11.8 | 5.00 | 1.5 | 0.010 | 0.01 | | | | 8 | 7.87 | 133 | | 1.79 |
| 169 | Eagle L | 6/12/2000 | 12.4 | 5.55 | 1.5 | 0.006 | 0.01 | | | | 7 | 7.57 | 129 | | 8.20 |
| 169 | Eagle L | 6/26/2000 | 11.5 | 7.95 | 1.5 | 0.010 | 0.01 | | | | 8 | 7.94 | 137 | | 0.56 |
| 169 | Eagle L | 7/10/2000 | 11.6 | 5.95 | 1.5 | 0.004 | 0.01 | | | | 3 | 7.49 | 139 | | 1.08 |
| 169 | Eagle L | 7/24/2000 | 11.5 | 7.00 | 1.5 | 0.004 | 0.01 | | | | 7 | 7.47 | 136 | | 1.15 |
| 169 | Eagle L | 8/7/2000 | 11.5 | 7.40 | 1.5 | 0.005 | 0.01 | | | | 4 | 6.82 | 137 | | 2.08 |
| 169 | Eagle L | 8/22/2000 | 11.7 | 7.00 | 1.5 | 0.008 | 0.01 | | | | 3 | 8.15 | 134 | | 0.42 |
| 169 | Eagle L | 9/4/2000 | 11.5 | 6.25 | 1.5 | 0.006 | 0.01 | | | | 6 | 6.75 | 138 | | 1.41 |
| 169 | Eagle L | 7/8/2001 | 11.5 | 7.00 | 1.5 | 0.006 | 0.01 | | | | 3 | 7.91 | 136 | | 1.51 |
| 169 | Eagle L | 7/22/2001 | 11.5 | 7.25 | 1.5 | 0.005 | 0.01 | | | | 4 | 7.94 | 138 | | 0.77 |
| 169 | Eagle L | 8/5/2001 | 11.5 | 7.35 | 1.5 | 0.005 | 0.01 | | | | 3 | 7.19 | 117 | | 0.99 |
| 169 | Eagle L | 9/3/2001 | 11.5 | 6.95 | 1.5 | 0.006 | 0.01 | | | | 4 | 7.64 | 139 | | |
| 169 | Eagle L | 9/30/2001 | 11.5 | 7.60 | 1.5 | 0.004 | 0.01 | | | | 6 | 6.80 | 142 | | |
| 169 | Eagle L | 10/10/2001 | 11.5 | 7.35 | 1.5 | 0.012 | 0.01 | | | | 5 | 7.57 | 145 | | |
| 169 | Eagle L | 10/23/2001 | 11.5 | 7.20 | 1.5 | 0.007 | 0.01 | | | | 8 | 7.31 | 153 | | 0.78 |
| 169 | Eagle L | 06/10/02 | 11.5 | 5.50 | 1.5 | 0.009 | 0.01 | 0.02 | 0.56 | 64.65 | 9 | 7.04 | 142 | | 1.04 |
| 169 | Eagle L | 06/23/02 | 11.5 | 6.50 | 1.5 | 0.009 | 0.00 | 0.03 | 0.40 | 46.00 | 2 | 7.70 | 140 | | |
| 169 | Eagle L | 07/07/02 | 11.5 | 5.90 | 1.5 | 0.007 | 0.00 | 0.02 | 0.38 | 52.77 | 11 | 7.93 | 141 | | 1.23 |
| 169 | Eagle L | 07/19/02 | 11.5 | 6.10 | 1.5 | 0.007 | 0.00 | 0.05 | 0.54 | 76.15 | 16 | 6.71 | 147 | | 0.75 |
| 169 | Eagle L | 07/21/02 | 11.5 | 6.95 | 1.5 | 0.006 | 0.00 | 0.03 | 0.37 | 66.32 | 8 | 7.63 | 143 | | |
| 169 | Eagle L | 08/04/02 | 11.5 | 6.30 | 1.5 | 0.006 | 0.00 | 0.04 | 0.74 | 128.35 | 6 | 7.95 | 144 | | 0.99 |
| 169 | Eagle L | 09/01/02 | 11.5 | 7.05 | 1.5 | 0.007 | 0.00 | 0.04 | 0.55 | 83.09 | 2 | 7.74 | 141 | | 1.11 |
| 169 | Eagle L | 09/21/02 | 11.5 | 6.85 | 1.5 | 0.002 | 0.00 | 0.01 | 0.33 | 172.97 | 5 | 7.94 | 142 | | 1.04 |
| 169 | Eagle L | 06/10/02 | 11.5 | | | 0.009 | 0.01 | 0.03 | 0.58 | 67.04 | | | | | |
| 169 | Eagle L | 06/23/02 | 11.5 | | | 0.009 | 0.00 | 0.01 | 0.31 | 34.03 | | | | | |
| 169 | Eagle L | 07/07/02 | 11.5 | | | 0.007 | 0.00 | 0.01 | 0.45 | 68.53 | | | | | |
| 169 | Eagle L | 07/19/02 | 11.5 | | | 0.004 | 0.00 | 0.03 | 0.51 | 121.37 | | | | | |
| 169 | Eagle L | 07/21/02 | 11.5 | | | 0.007 | 0.00 | 0.01 | 0.34 | 48.04 | | | | | |
| 169 | Eagle L | 08/04/02 | 11.5 | | | 0.006 | 0.00 | 0.04 | 0.51 | 83.32 | | | | | |
| 169 | Eagle L | 09/01/02 | 11.5 | | | 0.013 | 0.00 | 0.02 | 0.44 | 34.96 | | | | | |
| 169 | Eagle L | 09/21/02 | 11.5 | | | 0.002 | 0.00 | 0.01 | 0.33 | 214.16 | | | | | |

| LNum | PName | Date | TAir | TH20 | QA | QB | QC | QD |
|------|---------|------------|------|------|----|----|----|-----|
| 169 | Eagle L | 5/30/2000 | 20 | 15 | 2 | 1 | 3 | |
| 169 | Eagle L | 6/12/2000 | 17 | 16 | 2 | 2 | 3 | 25 |
| 169 | Eagle L | 6/26/2000 | 30 | 24 | 2 | 3 | 3 | 2 |
| 169 | Eagle L | 7/10/2000 | 22 | 22 | 2 | 3 | 3 | 56 |
| 169 | Eagle L | 7/24/2000 | 25 | 22 | 2 | 3 | 3 | 2 |
| 169 | Eagle L | 8/7/2000 | 21 | 23 | 2 | 3 | 3 | 125 |
| 169 | Eagle L | 8/22/2000 | 26 | 24 | 2 | 3 | 3 | 2 |
| 169 | Eagle L | 9/4/2000 | 11 | 21 | 2 | 3 | 3 | 25 |
| 169 | Eagle L | 7/8/2001 | 24 | 21 | 2 | 3 | 3 | 5 |
| 169 | Eagle L | 7/22/2001 | 24 | 24 | | | | |
| 169 | Eagle L | 8/5/2001 | 29 | 25 | | | | |
| 169 | Eagle L | 9/3/2001 | 27 | 24 | 2 | 3 | 3 | 2 |
| 169 | Eagle L | 9/30/2001 | 20 | 18 | | | | |
| 169 | Eagle L | 10/10/2001 | 11 | 15 | | | | |
| 169 | Eagle L | 10/23/2001 | 13 | 14 | 2 | 3 | 3 | 5 |
| 169 | Eagle L | 06/10/02 | 22 | 16 | 2 | 2 | 3 | 2 |
| 169 | Eagle L | 06/23/02 | 24 | 20 | 2 | 2 | 2 | 5 |
| 169 | Eagle L | 07/07/02 | 22 | 24 | 2 | 3 | 3 | 2 |
| 169 | Eagle L | 07/19/02 | 21 | 25 | 2 | 3 | 3 | 2 |
| 169 | Eagle L | 07/21/02 | 29 | 25 | 2 | 3 | 3 | 2 |
| 169 | Eagle L | 08/04/02 | 26 | 24 | 2 | 3 | 3 | 2 |
| 169 | Eagle L | 09/01/02 | 20 | 22 | 2 | 3 | 3 | 258 |
| 169 | Eagle L | 09/21/02 | 25 | 21 | 2 | 3 | 3 | 2 |
| 169 | Eagle L | 06/10/02 | | | | | | |
| 169 | Eagle L | 06/23/02 | | | | | | |
| 169 | Eagle L | 07/07/02 | | | | | | |
| 169 | Eagle L | 07/19/02 | | | | | | |
| 169 | Eagle L | 07/21/02 | | | | | | |
| 169 | Eagle L | 08/04/02 | | | | | | |
| 169 | Eagle L | 09/01/02 | | | | | | |
| 169 | Eagle L | 09/21/02 | | | | | | |

Appendix B. New York State Water Clarity Classifications

| | |
|-------------------------------|---|
| Class N: | Enjoyment of water in its natural condition and where compatible, as source of water for drinking or culinary purposes, bathing, fishing and fish propagation, recreation and any other usages except for the discharge of sewage, industrial wastes or other wastes or any sewage or waste effluent not having filtration resulting from at least 200 feet of lateral travel through unconsolidated earth. These waters should contain no deleterious substances, hydrocarbons or substances that would contribute to eutrophication, nor shall they receive surface runoff containing any such substance. |
| Class AA _{special} : | Source of water supply for drinking, culinary or food processing purposes; primary and secondary contact recreation; and fishing. These waters shall be suitable for fish propagation and survival, and shall contain no floating solids, settleable solids, oils, sludge deposits, toxic wastes, deleterious substances, colored or other wastes or heated liquids attributable to sewage, industrial wastes or other wastes. There shall be no discharge or disposal of sewage, industrial wastes or other wastes into these waters. These waters shall contain no phosphorus and nitrogen in amounts that will result in growths of algae, weeds and slimes that will impair the waters for their best usages. |
| Class A _{special} : | Source of water supply for drinking, culinary or food processing purposes; primary and secondary contact recreation; and fishing. These waters shall be suitable for fish propagation and survival. These international boundary waters, if subjected to approved treatment equal to coagulation, sedimentation, filtration and disinfection, with additional treatment if necessary to remove naturally present impurities, will meet New York State Department of Health drinking water standards and will be considered safe and satisfactory for drinking water purposes |
| Class AA: | Source of water supply for drinking, culinary or food processing purposes; primary and secondary contact recreation; and fishing. These waters shall be suitable for fish propagation and survival. These waters, if subjected to approved disinfection treatment, with additional treatment if necessary to remove naturally present impurities, will meet New York State Department of Health drinking water standards and will be considered safe and satisfactory for drinking water purposes |
| Class A: | Source of water supply for drinking, culinary or food processing purposes; primary and secondary contact recreation; and fishing. These waters shall be suitable for fish propagation and survival. These waters, if subjected to approved treatment equal to coagulation, sedimentation, filtration and disinfection, with additional treatment if necessary to remove naturally present impurities, will meet New York State Department of Health drinking water standards and will be considered safe and satisfactory for drinking water purposes |

- Class B Suitable for primary and secondary contact recreation and fishing. These waters shall be suitable for fish propagation and survival
- Class C: Suitable for fishing, and fish propagation and survival. The water quality shall be suitable for primary and secondary contact recreation, although other factors may limit the use for these purposes.
- Class D: Suitable for fishing. Due to such natural conditions as intermittency of flow, water conditions not conducive to propagation of game fishery, or stream bed conditions, the waters will not support fish propagation. These waters shall be suitable for fish survival. The water quality shall be suitable for primary and secondary contact recreation, although other factors may limit the use for these purposes.
- Class (T): Designated for trout survival, defined by the Environmental Conservation Law Article 11 (NYS, 1984b) as brook trout, brown trout, red throat trout, rainbow trout, and splake

APPENDIX C: BACKGROUND INFO FOR EAGLE LAKE

| | |
|---|---|
| CSLAP Number | 169 |
| Lake Name | Eagle L |
| First CSLAP Year | 2000 |
| Sampled in 2002? | yes |
| Latitude | 435218 |
| Longitude | 733702 |
| Elevation (m) | 288 |
| Area (ha) | 170.9 |
| Volume Code | 5 |
| Volume Code Name | Upper Hudson River |
| Pond Number | 438 |
| Qualifier | none |
| Water Quality Classification | B |
| County | Essex |
| Town | Ticonderoga |
| Watershed Area (ha) | not yet determined |
| Retention Time (years) | not yet determined |
| Mean Depth (m) | not yet determined |
| Runoff (m/yr) | 0.510947047 |
| Watershed Number | 11 |
| Watershed Name | Upper Hudson River |
| NOAA Section | 3 |
| Closest NOAA Station | North Creek |
| Closest USGS Gaging Station-Number | 4276842 |
| Closest USGS Gaging Station-Name | Putnam Point East of Crown Point Center |
| CSLAP Lakes in Watershed | Adirondack L, Babcock L, Ballston L, Brant L, Cossayuna L, Eagle L, Efner L, Friends L, Garnet L, Goodnow F, Hedges L, Hunt L, Kellum L, L Forest, L Lauderdale, L Luzerne, Loon L-W, Mayfield L, Moreau L, Piseco L, Sacandaga L, Saratoga L, Schroon L, Summit L-W, Taconic P, Windover L |